

1 ***In vitro* assessment of the impact of nickel on the viability and steroidogenesis in**  
2 **the human adrenocortical carcinoma (NCI-H295R) cell line**

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30 **Short title:** Nickel in Relation to Sexual Steroid Hormones

31

32 **Summary**

33 Nickel is a ubiquitous environmental pollutant, which has various effects on  
34 reproductive endocrinology. In this study, human adrenocortical carcinoma (NCI-  
35 H295R) cell line was used as an *in vitro* biological model to study the effect of nickel  
36 chloride (NiCl<sub>2</sub>) on the viability and steroidogenesis. The cells were exposed to  
37 different concentrations (3.90; 7.80; 15.60; 31.20; 62.50; 125; 250 and 500 μM) of  
38 NiCl<sub>2</sub> and compared with control group (culture medium without NiCl<sub>2</sub>). The cell  
39 viability was measured by the metabolic activity assay. Production of sexual steroid  
40 hormones was quantified by enzyme linked immunosorbent assay. Following 48 h  
41 culture of the cells in the presence of NiCl<sub>2</sub> a dose-dependent depletion of progesterone  
42 release was observed even at the lower concentrations. In fact, lower levels of  
43 progesterone were detected in groups with higher doses ( $\geq 125 \mu\text{M}$ ) of NiCl<sub>2</sub> ( $P < 0.01$ ),  
44 which also elicited cytotoxic action. A more prominent decrease in testosterone  
45 production ( $P < 0.01$ ) was also noted in comparison to that of progesterone. On the  
46 other hand, the release of 17 $\beta$ -estradiol was substantially increased at low  
47 concentrations (3.90 to 62.50 μM) of NiCl<sub>2</sub>. The cell viability remained relatively  
48 unaltered up to 125 μM ( $P > 0.05$ ) and slightly decreased from 250 μM of NiCl<sub>2</sub> ( $P <$   
49 0.05). Our results indicate endocrine disruptive effect of NiCl<sub>2</sub> on the release of  
50 progesterone and testosterone in the NCI-H295R cell line. Although no detrimental  
51 effect of NiCl<sub>2</sub> ( $\leq 62.50 \mu\text{M}$ ) could be found on 17 $\beta$ -estradiol production, its toxicity  
52 may reflect at other points of the steroidogenic pathway.

53

54 **Keywords:** nickel chloride, sexual steroid hormones, cell viability, NCI-H295R cell  
55 line, endocrine disruption

56

## 57 **Introduction**

58

59 Several environmental contaminants are recognized as endocrine disruptors (EDs),  
60 which may adversely affect the reproductive functions of humans, as well as wildlife  
61 species (Kabir *et al.* 2015, Vitku *et al.* 2015, Yang *et al.* 2015, Roychoudhury *et al.*  
62 2016, Kolatorova *et al.* 2017, Jambor *et al.* 2018, Jambor *et al.* 2019). This  
63 heterogeneous group of exogenous substances has the ability to alter functions of the  
64 endocrine system with a subsequent negative impact on the cellular behaviour and  
65 health in an intact organism. Endocrine disruptors may be found in a variety of  
66 products, such as pesticides, household items, cosmetics or plastic packaging. It is likely  
67 that some EDs are structural analogues of steroids, having similar effects as true  
68 hormones, high levels of which may have disproportionate consequences (Sanderson  
69 2006, Svechnikov *et al.* 2010). They can strongly affect reproductive and endocrine  
70 functions in several ways (Andersen *et al.* 2002), either by directly affecting the  
71 hormone production through interaction with the appropriate enzymes, or through  
72 interfering with their transport to target organs to alter natural hormone metabolism or  
73 even to inactivate the function of steroidogenesis regulatory proteins (e.g.,  
74 Steroidogenic Acute Regulatory – StAR) (Sanderson and van den Berg 2003).

75 Nickel (Ni) is a widely distributed metal that is industrially applied in various mineral  
76 forms (Lu *et al.* 2005). Dusts from volcanic emissions, the weathering of rocks and  
77 soils, biological cycles and solubilisation of Ni compounds from soils (Sunderman  
78 2004) represent the main natural sources of atmospheric (7.0-12.0 ng of Ni/m<sup>3</sup> of air;

79 150 ng of Ni/m<sup>3</sup> of air near point sources) and aqueous (3-10 µg of Ni/l of water in  
80 surface water and groundwater) Ni (ATSDR 2005). Anthropogenic sources of Ni  
81 pollution include mining, smelting and refining activities, burning of fossil fuels,  
82 sewage incineration and plastic production (Yu 2005). The major source of exposure to  
83 Ni for the general population is the food chain (Pandey and Srivastava 2000, Llamas  
84 and Sanz 2008). It usually enters the body via food and water consumption, although  
85 inhalation exposure in occupational settings is the primary route for Ni-induced toxicity  
86 (Ankel-Fuchs and Thauer 1988). Based on the average levels of Ni consumption  
87 through water (4.0-8.6 µg/day) and food (69.0-162.0 µg/day), the daily per oral intake  
88 of the metal was estimated to be 0.001-0.0024 mg/kg/day for an average adult human  
89 being weighing 70 kg (ATSDR 2005). Another source of non-occupational exposure to  
90 Ni is tobacco smoking, and each cigarette is estimated to contain 1.1-3.1 µg of Ni  
91 (Cempel and Nickel 2006).

92 Although Ni is considered to be an essential micronutrient (Eisler 1998), it has a  
93 number of effects in the cell (Das 2009). It plays an important role in DNA, RNA and  
94 protein structure and/or function (Pandey and Srivastava 2000). Nickel also serves as a  
95 cofactor or a structural component of several metalloenzymes (Przybyla *et al.* 1992).  
96 Deficiency is rare due to a low level of requirement, and relatively high availability in  
97 the diet, but experiments have shown that at cellular levels Ni deprivation may result in  
98 changes in the membrane properties and other structures (Das and Dasgupta 1997, Das  
99 2009). On the contrary, high quantity of Ni is injurious for animal and human health  
100 (Pandey *et al.* 1999, Pandey and Srivastava 2000). More recently, several reports have  
101 showed that Ni is able to induce toxicological, physiological and histopathological  
102 alterations in a number of animal species (Pane *et al.* 2003, Bersenyi *et al.* 2004, Brix *et*  
103 *al.* 2004, Gupta *et al.* 2006, Krockova *et al.* 2011, Lukac *et al.* 2011). Soluble Ni  
104 compounds are likely to be human carcinogens (Costa 1991, Costa *et al.* 2005), and

105 toxic and/or carcinogenic effects of such Ni compounds may be associated with Ni-  
106 mediated oxidative damage to DNA, proteins and inhibition of cellular antioxidant  
107 defences (Rodriguez *et al.* 1996). There is sufficient evidence that Ni ions ( $\text{Ni}^{2+}$ ) have  
108 potential toxic effects on the reproductive system (Das and Dasgupta 2000). Animal  
109 studies referred to the negative effects of  $\text{Ni}^{2+}$  on the structure and function of testis,  
110 seminal vesicles, prostate gland (Pandey *et al.* 1999, Forgacs *et al.* 2001, Massanyi *et*  
111 *al.* 2003, Massanyi *et al.* 2007, Zemanova *et al.* 2007), and spermatozoa concentration  
112 as well as motility (Das and Dasgupta 2000, Lukac *et al.* 2011). Nickel salts are also  
113 capable of inducing morphological changes such as, degeneration of testicular germinal  
114 epithelium (Benson *et al.* 1988, Pandey *et al.* 1999), testicular sarcomas as well as  
115 functional disorders including inhibition of spermatogenesis (Mathur *et al.* 1977, Yokio  
116 *et al.* 2003) and steroidogenesis (Das and Dasgupta 2002, Krockova *et al.* 2011). Such  
117 negative effects may ultimately lead to sterility (Massanyi *et al.* 2007).

118 Steroidogenesis can be tested using a number of cell lines or primary culture with  
119 gonadal tissue, but the most widely used assay utilizes a human adrenocortical  
120 carcinoma (NCI-H295R) cell line. Such *in vitro* steroidogenesis screening assays are  
121 used to examine the impact of endocrine active chemicals/substances (EACs) capable of  
122 altering steroid biosynthesis (Ding *et al.* 2007, Fialkova *et al.* 2018). Progesterone,  
123 testosterone and estradiol are the main steroid hormones that play essential roles during  
124 the regulation of reproduction in vertebrates and are also involved in numerous other  
125 processes related to development and growth (Hecker and Giesy 2008). The present  
126 study investigated the effects of nickel chloride ( $\text{NiCl}_2$ ) on the viability and  
127 steroidogenesis of the NCI-H295R cell line. Specifically, we examined the dose-  
128 dependent changes of  $\text{NiCl}_2$  as a potential endocrine disruptor in relation to the release  
129 of progesterone, testosterone and  $17\beta$ -estradiol by NCI-H295R cell line *in vitro*. The  
130 NCI-H295R cell line was derived from H295 cells which were established from a

131 primary hormonally active adrenocortical carcinoma (Gazdar *et al.* 1990, Rainey *et al.*  
132 2004). This cell line has physiological characteristics of zonally undifferentiated human  
133 fetal adrenal cells (Staels *et al.* 1993, Harvey and Everett 2003), and represent an unique  
134 *in vitro* model system having the ability to produce all of the steroid hormones found in  
135 the adult adrenal cortex and the gonads, allowing testing the effects on both  
136 corticosteroid synthesis together with the production of sexual steroid hormones  
137 (Gazdar *et al.* 1990). Another advantage of the H295R cell bioassay is that it can be  
138 used to evaluate the enzymatic activities of steroidogenic genes (Hilscherova *et al.*  
139 2004). In fact, the NCI-H295R Steroidogenesis Assay has been included in the Tier1  
140 Screening Battery of the United States Environmental Protection Agency's (EPA)  
141 Endocrine Disruptor Screening Program (EDSP). The test guideline of the H295R  
142 Steroidogenesis Assay (TG 456) has been further validated by the Organization for  
143 Economic Cooperation and Development (OECD 2011).

144

## 145 **Materials and Methods**

146

### 147 *Cell culture*

148 The NCI-H295R cell line was obtained from the American Type Culture Collection  
149 (ATCC, Manassas, VA, USA). The cells were cultured in a Good Laboratory Practice  
150 (GLP) certified laboratory (National Institute of Chemical Safety, Budapest, Hungary;  
151 OGYI/31762-9/2010) according to previously established and specifically validated  
152 protocols (Hilscherova *et al.* 2004, Zhang *et al.* 2005, Hecker *et al.* 2006, Hecker and  
153 Giesy 2008, OECD 2011).

154 After initiation of the NCI-H295R culture from the original ATCC batch, cells were  
155 cultured for five passages and these cells were split and frozen down in liquid nitrogen  
156 (-196°C). The cells for the experiments were cultured for a minimum of five additional

157 passages using new NCI-H295R batches from frozen stocks prior to initiation of the  
158 exposure studies. The cells were grown in 75 cm<sup>2</sup> plastic cell culture flasks (TPP  
159 Techno Plastic Products AG, Switzerland) in an incubator under standard conditions  
160 (37°C and 5.0% CO<sub>2</sub> atmosphere). Subsequently, the cells were grown in a 1:1 mixture  
161 of Dulbecco's Modified Eagle's Medium and Ham's F-12 Nutrient mixture  
162 (DMEM/F12; Sigma-Aldrich, St. Louis, MO, USA) supplemented with 1.2 g/l NaHCO<sub>3</sub>  
163 (Sigma-Aldrich, St. Louis, MO, USA), 5.0 ml/l of ITS+Premix (BD Bioscience, San  
164 Jose, CA, USA) and 12.5 ml/l of BD Nu-Serum (BD Bioscience, San Jose, CA, USA).  
165 The medium was changed 2-3 times per week and cells were detached from flasks for  
166 sub-culturing using sterile 0.25% trypsin-EDTA (Sigma-Aldrich, St. Louis, MO, USA).  
167 After trypsinization, cells were plated at the appropriate density to obtain 90-100%  
168 confluency. Cell density was determined using a hemocytometer (Fig. 1) and adjusted  
169 with culture medium to a final concentration of 300 000 cells/ml. The cell suspensions  
170 were plated (with final volume of 1.0 ml/well) into sterile plastic 24-well plates (TPP,  
171 Grainer, Germany) for estimation of sexual steroid hormones (50-60% confluency of  
172 cells). For cytotoxicity evaluation, the cells (100 µl/well) were seeded into 96-well  
173 plates (MTP, Grainer, Germany). The seeded plates were incubated at 37°C and 5.0%  
174 CO<sub>2</sub> atmosphere for 24 h to allow the cells to attach to the wells (Knazicka *et al.* 2013).

175

#### 176 *In vitro exposure*

177 After 24 h attachment period, the cell culture medium was removed from the plates and  
178 replaced with a new medium supplemented with 3.90; 7.80; 15.60; 31.20; 62.50; 125;  
179 250 and 500 µM nickel chloride (NiCl<sub>2</sub>; ≥ 98%; Sigma-Aldrich, St. Louis, MO, USA),  
180 respectively. Cell cultures were set in 24 and 96-well plates (MTP, Grainer, Germany).  
181 Following treatment, the cells were maintained for 48 h. The experimental groups A - H

182 (exposed to different concentrations of NiCl<sub>2</sub>) with control group (Ctrl) (culture medium  
183 without NiCl<sub>2</sub>) were compared.

184

#### 185 *Cell viability*

186 The viability of the cells exposed to NiCl<sub>2</sub> was evaluated by the metabolic activity  
187 (MTT) assay (Mosmann 1983). This colorimetric assay measures the conversion of  
188 a yellow tetrazolium salt [3-(4,5-dimethylthiazol-2-yl)-2,5-diphenyltetrazolium bromide  
189 i.e. MTT), to blue formazan particles by mitochondrial succinate dehydrogenase  
190 enzyme of intact mitochondria of living cells. Formazan was measured  
191 spectrophotometrically. Following the termination of NiCl<sub>2</sub> exposure, the cells were  
192 stained with MTT (Sigma-Aldrich, St. Louis, MO, USA) at a final concentration of 0.2  
193 mg/ml. After 2 h incubation (37°C, and 5.0% CO<sub>2</sub> atmosphere), the cells and the  
194 formazan crystals were dissolved in 150 µl of acidified (0.08 M HCl) isopropanol  
195 (CentralChem, Bratislava, Slovak Republic). The absorbance was determined at  
196 a measuring wavelength of 570 nm against 620 nm as reference by a microplate reader  
197 (Anthos MultiRead 400, Austria). The data were expressed in percentage of the control  
198 group (i.e., absorbance of formazan from cells not exposed to NiCl<sub>2</sub>).

199

#### 200 *Hormonal analysis*

201 At the end of 48 h NiCl<sub>2</sub> exposure, the aliquots of the culture medium were removed  
202 from the 24-well cell culture plates and after centrifugation the supernatant was  
203 collected and frozen at -80°C until sexual steroid hormones measurements. Enzyme  
204 linked immunosorbent assay (ELISA) was used for the quantification of progesterone,  
205 testosterone and 17β-estradiol (Dialab GmbH, Wiener Neudorf, Austria) directly from  
206 the aliquots of the medium. According to the manufacturer's data, the sensitivity of  
207 testosterone assay was 0.075 ng/ml, and the intra- and inter-assay coefficients of



208 variation were 4.6% and 7.5%, respectively. Cross-reactivity with 5 $\alpha$ -  
209 dihydrotestosterone was 16.0%. The sensitivity of progesterone assay was 0.05 ng/ml,  
210 and the intra- and inter-assay coefficients of variation were  $\leq$  4.0% and  $\leq$  9.3%,  
211 respectively. The intra- and inter-assay coefficients of variation for the 17 $\beta$ -estradiol  
212 assay were  $\leq$  9.0% and  $\leq$  10.0%, and the sensitivity was 8.68 pg/ml. The absorbance  
213 was determined at a wavelength 450 nm using a microplate reader (Anthos MultiRead  
214 400, Austria) and the data were evaluated by WinRead 2.30 computer software. Values  
215 were expressed in percentage of the untreated control (control groups served as 100%).  
216 Forscolin, prochloraz and aminoglutethimide (Sigma-Aldrich, St. Louis, MO, USA)  
217 dissolved in 0.1% DMSO were used as positive controls.

218

#### 219 *Statistical analysis*

220 Obtained data were statistically analyzed using the PC program GraphPad Prism 3.02  
221 (GraphPad Software Incorporated, San Diego, California, USA). Descriptive statistical  
222 characteristics (arithmetic mean, minimum, maximum, standard deviation and  
223 coefficient of variation) were evaluated. Homogeneity of variance was assessed by  
224 Bartlett's test. One-way analysis of variance (ANOVA) and the Dunnett's multiple  
225 comparison tests were used for statistical evaluations. The level of significance was set  
226 at \*\*\* (P < 0.001); \*\* (P < 0.01) and \* (P < 0.05). Three independent experiments were  
227 performed.

228

## 229 **Results**

230

### 231 *Cell viability*

232 The cell viability remained relatively unaltered up to 125  $\mu$ M (P > 0.05) and slightly  
233 decreased from 250  $\mu$ M of NiCl<sub>2</sub> (P < 0.05). The cytotoxic effect of NiCl<sub>2</sub> (< 50%) was

234 very distinct ( $P < 0.01$ ) in the group with the highest concentration (500  $\mu\text{M}$ ) of  $\text{NiCl}_2$   
 235 (Fig. 2).

236

237 *Release of progesterone by human adrenocortical carcinoma (NCI-H295R) cell line*

238 Following 48 h culture of NCI-H295R cell line in the presence of  $\text{NiCl}_2$ , a dose-  
 239 dependent depletion ( $P < 0.01$ ) of progesterone release was observed in all the  
 240 experimental groups, even at the lowest concentration (3.90  $\mu\text{M}$ ) of  $\text{NiCl}_2$  used in the  
 241 study ( $19.56 \pm 4.00$  ng/ml). Lower levels of progesterone were detected in groups with  
 242 higher doses ( $\geq 125$   $\mu\text{M}$ ) of  $\text{NiCl}_2$  ( $P < 0.01$ ) as shown in Table 1. In the control group,  
 243 progesterone production (100%) was  $21.05 \pm 4.40$  ng/ml. The percentage changes of  
 244 progesterone release after  $\text{NiCl}_2$  exposure are presented in Fig. 3.

245

246 **Table 1.** Effect of 48 h  $\text{NiCl}_2$  exposure on the release of progesterone (ng/ml) by human  
 247 adrenocortical carcinoma (NCI-H295R) cell line. *Abbreviations:* x – arithmetic mean,  
 248  $\pm\text{S.D.}$  – standard deviation, CV (%) – coefficient of variation. The level of significance  
 249 was set at \*\*\* ( $P < 0.001$ ), \*\* ( $P < 0.01$ ) and \* ( $P < 0.05$ ). Ctrl – control group.

250

	Control	3.90	7.80	15.60	31.20	62.50	125	250	500
Groups	Ctrl	H	G	F	E	D	C	B	A
	NiCl <sub>2</sub> ( $\mu\text{M}$ )								
<b>X</b>	21.05	19.56**	11.07**	10.59**	11.70**	9.93**	7.33**	6.18**	5.79**
<b>minimum</b>	15.24	15.64	6.24	6.84	8.27	6.88	5.27	4.25	3.55
<b>maximum</b>	28.25	24.12	14.25	15.26	14.85	14.49	10.58	8.18	7.58
<b><math>\pm\text{S.D.}</math></b>	4.40	4.00	3.26	3.06	2.65	3.00	1.99	1.62	1.35
<b>CV (%)</b>	20.91	20.46	29.48	28.96	22.96	30.24	27.15	26.13	23.41
<b>%</b>	100.00	92.92	52.59	47.51	55.58	47.19	34.83	29.37	27.48

251

252 *Release of testosterone by human adrenocortical carcinoma (NCI-H295R) cell line*

253 Testosterone production decreased significantly ( $P < 0.01$ ) at all the concentrations of  
 254  $\text{NiCl}_2$  used in the study (Table 2). Furthermore, this decline was more prominent in  
 255 comparison to that of progesterone. The lowest release of testosterone was ( $P < 0.01$ )  
 256 noted at 125  $\mu\text{M}$  of  $\text{NiCl}_2$  ( $1.22 \pm 0.74$  ng/ml) in comparison with control group ( $10.75$   
 257  $\pm 3.45$  ng/ml). The percentage changes of testosterone release after  $\text{NiCl}_2$  exposure are  
 258 presented in Fig. 4.

259

260 **Table 2.** Effect of 48 h  $\text{NiCl}_2$  exposure on the release of testosterone (ng/ml) by human  
 261 adrenocortical carcinoma (NCI-H295R) cell line. *Abbreviations:* x – arithmetic mean,  
 262  $\pm$ S.D. – standard deviation, CV (%) – coefficient of variation. The level of significance  
 263 was set at \*\*\* ( $P < 0.001$ ), \*\* ( $P < 0.01$ ) and \* ( $P < 0.05$ ). Ctrl – control group.

264

	Control	3.90	7.80	15.60	31.20	62.50	125	250	500
Groups	Ctrl	H	G	F	E	D	C	B	A
	NiCl <sub>2</sub> ( $\mu\text{M}$ )								
<b>X</b>	10.75	4.42**	3.18**	1.98**	4.96**	1.46**	1.22**	2.18**	1.86**
<b>minimum</b>	6.54	2.12	1.84	0.48	3.02	0.27	0.25	0.88	0.57
<b>maximum</b>	16.44	7.28	5.87	3.54	7.12	2.71	2.33	3.19	3.21
<b><math>\pm</math>S.D.</b>	3.45	2.02	1.47	1.18	1.71	0.90	0.74	0.89	1.07
<b>CV (%)</b>	32.12	45.66	46.33	59.52	34.45	61.85	60.43	41.05	57.51
<b>%</b>	100.00	41.10	29.55	18.41	46.14	13.53	11.30	20.23	17.26

265

266 *Release of 17 $\beta$ -estradiol by human adrenocortical carcinoma (NCI-H295R) cell line*

267 The 17 $\beta$ -estradiol production was substantially increased at low concentrations (3.90 to  
 268 62.50  $\mu\text{M}$ ) of  $\text{NiCl}_2$ . However, the increment was not statistically significant ( $P > 0.05$ )  
 269 in comparison with control group (Fig. 5). The lowest release of 17 $\beta$ -estradiol by NCI-

270 H295R cell line was recorded in groups with high concentrations ( $\geq 125 \mu\text{M}$ ) of  $\text{NiCl}_2$ ,  
 271 which released similar levels of  $17\beta$ -estradiol (Table 3).

272

273 **Table 3.** Effect of 48 h  $\text{NiCl}_2$  exposure on the release of  $17\beta$ -estradiol (pg/ml) by human  
 274 adrenocortical carcinoma (NCI-H295R) cell line. *Abbreviations:* x – arithmetic mean,  
 275  $\pm$ S.D. – standard deviation, CV (%) – coefficient of variation. The level of significance  
 276 was set at \*\*\* ( $P < 0.001$ ), \*\* ( $P < 0.01$ ) and \* ( $P < 0.05$ ). Ctrl – control group.

277

	Control	3.90	7.80	15.60	31.20	62.50	125	250	500
Groups	Ctrl	H	G	F	E	D	C	B	A
	$\text{NiCl}_2$ ( $\mu\text{M}$ )								
<b>X</b>	1.10	1.69	2.08	1.66	1.42	1.34	0.89	0.86	0.91
<b>Minimum</b>	0.74	0.94	1.02	0.80	0.84	0.94	0.64	0.76	0.51
<b>Maximum</b>	1.68	2.88	2.57	2.39	2.78	2.51	1.30	1.05	1.06
<b><math>\pm</math>S.D.</b>	0.29	0.67	0.57	0.68	0.71	0.60	0.20	0.09	0.18
<b>CV (%)</b>	26.32	39.86	27.57	41.23	49.90	45.05	21.87	10.49	20.35
<b>%</b>	100.00	154.70	167.00	151.70	129.60	122.40	81.73	78.37	83.21

278

279

## 280 Discussion

281

282 Hormonal effects are believed to play an important role in the reproductive toxicology  
 283 of Ni at both the neuroendocrine and gonadal levels in the hypothalamic-pituitary-  
 284 gonadal (HPG) axis (Forgacs *et al.* 2012). The effects of Ni on steroidogenesis have  
 285 been described recently; however, the results vary depending on the experimental  
 286 model, duration of exposure as well as the doses used. The present study on the impact  
 287 of  $\text{NiCl}_2$  on the NCI-H295R cell line suggests a direct action of  $\text{NiCl}_2$  on the steroid-

288 producing cells and subsequent changes in hormonal release. Nickel significantly  
289 decreased the release of progesterone and testosterone in the entire range of  
290 concentrations of NiCl<sub>2</sub> used in the study whereas the cell viability remained relatively  
291 unaltered up to 125 μM (P > 0.05) and slightly decreased from 250 μM of NiCl<sub>2</sub> (P <  
292 0.05). The cytotoxic effect of NiCl<sub>2</sub> (< 50%) was evident (P < 0.01) in the group with  
293 the highest concentration (500 μM) of NiCl<sub>2</sub>. These results clearly confirm reports of  
294 Forgacs *et al.* (2011) and Ocztos *et al.* (2011), who observed similar effects of Ni<sup>2+</sup>,  
295 Hg<sup>2+</sup> and Cd<sup>2+</sup> on the release of progesterone and testosterone by NCI-H295R cell line.  
296 Using primary gonadal culture, these authors also confirmed that Ni<sup>2+</sup> is able to disturb  
297 the sexual steroid production far below its cytotoxic concentration. Similar effects of  
298 other metals (cadmium, mercury, copper) have also been reported by our group from  
299 studies in the NCI-H295R cell line (Knazicka *et al.* 2013, 2015, Bilcikova *et al.* 2020).  
300 Earlier, Krockova and Massanyi (2010) reported a dose-dependent decrease in  
301 progesterone production by the Leydig cells at the highest concentration of 1000 μmol/l  
302 of NiCl<sub>2</sub>. Revesz *et al.* (2004) previously exposed human ovarian granulosa cells  
303 (obtained from women undergoing *in vitro* fertilization) to 15.60 to 1000 μM of Ni<sup>2+</sup> for  
304 48 h in order to determine the site of action of Ni<sup>2+</sup>. The granulosa cells were stimulated  
305 to produce progesterone by using maximally stimulating amounts of human chorionic  
306 gonadotropin (0.10 IU/ml hCG) or dibutyryl cyclic adenosine monophosphate (1.00  
307 mM db-cAMP). Dose-dependent depression in both hCG and db-cAMP stimulated  
308 progesterone production was seen at 15.60 μM or higher concentration of Ni<sup>2+</sup> which is  
309 not cytotoxic to human ovarian granulosa cells. The viability of cells remained  
310 unaffected up to 31.25 μM of Ni<sup>2+</sup> and decreased significantly at 62.50 μM of Ni<sup>2+</sup>.  
311 Their data further indicated that the effect of Ni<sup>2+</sup> on the progesterone production is not  
312 due to cytotoxicity, and the cellular site(s) of inhibitory action appears to be subsequent  
313 to the membrane receptor and production of db-cAMP. The inhibition of progesterone

314 secretion by granulosa cells (Roychoudhury *et al.* 2014a, 2015) or rat ovarian fragments  
315 (Roychoudhury *et al.* 2014b) were also induced by other metals. The effect of Ni<sup>2+</sup> may  
316 be associated with its interactions with other essential divalent metal cations, blocking  
317 functional groups, displacing essential metal ions or modifying active conformation of  
318 biomolecules (Coogan *et al.* 1989). Ni<sup>2+</sup> is known to inhibit calcium (Ca<sup>2+</sup>) channels.  
319 On the other hand, Ca<sup>2+</sup> plays an important role in the regulation of progesterone  
320 production as shown in the rat granulosa cells (Tsang and Carnegie 1983). In addition,  
321 Ni has been demonstrated to alter the metabolic activity of microsomal monooxygenases,  
322 some of which are essential for steroid metabolism (Mattison *et al.* 1983). Thus, above  
323 mentioned findings could also participate in Ni-triggered alterations of progesterone  
324 release by NCI-H295R cell line.

325 Our presented data showed that testosterone seemed to be more vulnerable than  
326 progesterone and 17 $\beta$ -estradiol to NiCl<sub>2</sub> exposure suggesting multiple sites of action of  
327 this metal in steroidogenesis. Disorders of the testosterone synthesis could result in a  
328 reduced activity of the key enzymes involved in the biosynthesis of testosterone. Das  
329 and Dasgupta (2002) reported that nickel sulphate (NiSO<sub>4</sub>) affects steroidogenic  
330 enzymes (3 $\beta$ -hydroxysteroid dehydrogenase and 17 $\beta$ -hydroxysteroid dehydrogenase)  
331 causing alterations in the testosterone formation in adult rat testes. In another study,  
332 Krockova *et al.* (2011) investigated the effects of NiCl<sub>2</sub> on the testosterone secretion,  
333 cell viability and apoptosis in mouse Leydig cells *in vitro*. They demonstrated that  
334 NiCl<sub>2</sub> decreased the testosterone production at a low dose (15.67  $\mu$ mol/l) and  
335 subsequently confirmed Ni-induced structural and functional alterations in the Leydig  
336 cells. Testosterone production by mouse primary Leydig cells culture following an *in*  
337 *vitro* Ni<sup>2+</sup> exposure (62.50 to 1000  $\mu$ M) was also evaluated by Forgacs *et al.* (1998).  
338 Dose-dependent depression in hCG-stimulated testosterone production was found at  $\geq$   
339 125  $\mu$ M or higher dose of Ni<sup>2+</sup>, while basal testosterone production remained

340 unaffected. They further showed the effect to be dose-dependent, and is not due to  
341 cytotoxicity. Previously, Laskey and Phelps (1991) examined the effect of Ni<sup>2+</sup> and  
342 other metal cations (Co<sup>2+</sup>, Cu<sup>2+</sup>, Hg<sup>2+</sup>, Cd<sup>2+</sup> and Zn<sup>2+</sup>) on *in vitro* Leydig cell  
343 testosterone production. The results showed no change in Leydig cell viability with any  
344 metal cation treatment during the 3 h incubation. Dose-response depression in both  
345 hCG- and db-cAMP-stimulated testosterone production was noted with Cd<sup>2+</sup>, Co<sup>2+</sup>,  
346 Cu<sup>2+</sup>, Hg<sup>2+</sup>, Ni<sup>2+</sup> and Zn<sup>2+</sup> treatment. Surprisingly, Cd<sup>2+</sup>, Co<sup>2+</sup>, Ni<sup>2+</sup> and Zn<sup>2+</sup> caused a  
347 depletion in hCG- and db-cAMP-stimulated testosterone production, also caused  
348 significant increases in 20 $\alpha$ -hydroxycholesterol- and pregnenolone-stimulated  
349 testosterone production over untreated and similarly stimulated cultures. This indicates  
350 that these cations may act at multiple sites within the Leydig cells. Sun *et al.* (2003)  
351 studied the mechanisms of changes in the genital system caused by nickel sulfate  
352 (NiSO<sub>4</sub>) in male rats. They observed that the contents of testicular Ni were increased;  
353 however, the blood serum contents of testosterone, follicle-stimulating hormone (FSH)  
354 and luteinizing hormone (LH) were reduced. It was assumed that the Ni-induced genital  
355 system injury in male rats may be related to the decrease in the content of these  
356 hormones.

357 The present study noted that the 17 $\beta$ -estradiol production was increased (although non-  
358 significantly) at low concentrations (3.90 to 62.50  $\mu$ M) of NiCl<sub>2</sub> (P > 0.05). In  
359 agreement with our results, no significant changes were observed in serum estradiol  
360 levels in rats intraperitoneally injected with NiCl<sub>2</sub> (4 mg/kg body weight) (Hfaiedh *et al.*  
361 2007). In the treated rats, demonstrably increased activity of testicular aromatase was  
362 also reported. Taking into account these considerations we presume that the  
363 considerably decreased levels of testosterone together with non-significant alterations in  
364 release of 17 $\beta$ -estradiol in the present study could be associated with higher aromatase

365 activity leading to stable estrogen levels as an adaptive response of NCI-H295R cell line  
366 to Ni exposure.

367 As a metalloestrogen, Ni activates estrogen receptor- $\alpha$  (ER $\alpha$ ) (Darbre 2006, Forgacs *et*  
368 *al.* 2012). Martin *et al.* (2003) examined the ability of metal ions to activate ER $\alpha$  in the  
369 human breast cancer cell line (MCF-7). Similar to estradiol, treatment of cells with Cu,  
370 Co, Ni, Pb, Hg, Sn, Cr or V stimulated cell proliferation. The metals also decreased the  
371 concentration of ER $\alpha$  protein and mRNA, and induced expression of the estrogen-  
372 regulated genes, progesterone receptor and pS2. The ability of such metals to alter gene  
373 expression was blocked by an anti-estrogen, suggesting that their activity is probably  
374 mediated by ER $\alpha$ . The estrogenic potency of Ni was comparable to that of estradiol.

375 Moreover, the cytotoxic effect of NiCl<sub>2</sub> (< 50%) was evident (P < 0.01) in the group  
376 with the highest concentration (500  $\mu$ M/ml) of NiCl<sub>2</sub> used in the study. The cell  
377 viability remained relatively unaltered up to 125  $\mu$ M (P > 0.05) and slightly decreased  
378 from 250  $\mu$ M of NiCl<sub>2</sub> (P < 0.05). Ng and Liu (1990) noted that Ni (1.0; 10.0 and 100  
379  $\mu$ M of NiCl<sub>2</sub>.6H<sub>2</sub>O) and other metals tested (including PbCl<sub>2</sub>, ZnCl<sub>2</sub>, AlCl<sub>3</sub>, CrCl<sub>3</sub>,  
380 FeCl<sub>2</sub> and LiCl) had no deleterious effect on viability and hormone-induced  
381 steroidogenesis of Leydig cells and the cells in the adrenal gland.

382

### 383 **Conclusion**

384

385 The results of the present study indicate the endocrine disruptive effect of NiCl<sub>2</sub> on the  
386 release of sexual steroid hormones (progesterone and testosterone) in the human  
387 adrenocortical carcinoma (NCI-H295R) cell line even at low (minimum) concentrations.  
388 Testosterone release seemed more vulnerable whereas no detrimental effect of NiCl<sub>2</sub>  
389 could be seen at concentrations  $\leq$  62.50  $\mu$ M of NiCl<sub>2</sub> on 17 $\beta$ -estradiol production  
390 thereby suggesting multiple sites of action of this metal in the steroidogenic pathway.



391 Further research may clarify the precise molecular mechanism of action of NiCl<sub>2</sub> on the  
392 sexual steroid production and their metabolites whose production is conditioned by the  
393 steroidogenic enzymes.

394

#### 395 **Conflict of interest**

396 There is no conflict of interest.

397

#### 398 **Acknowledgments**

399 This study was financially supported by the Scientific Agency of the Slovak Republic  
400 VEGA No. 1/0163/18, APVV-15-0543, APVV-16-0289 and co-funded by European  
401 Community under project No. 26220220180: Building Research Centre  
402 „AgroBioTech”.

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644 **Figure 1.** Monolayer of human adrenocortical carcinoma (NCI-H295R) cell line.

645 *Abbreviations:* A – low density; B – high density (magnification 100x).

646

647 **Figure 2.** The viability of NCI-H295R cell line in culture after 48 h of NiCl<sub>2</sub> exposure.

648 *Abbreviations:* The cytotoxicity was assessed using the MTT assay following NiCl<sub>2</sub>

649 exposure. Each point represents the arithmetic mean ( $\pm$ S.D.) absorbance in % of

650 (untreated) controls (Ctrl) determined in three independent experiments. The number of

651 replicate wells was 22-32 at each point. A decline in absorbance reflects a decline in cell

652 viability. The statistical difference between the values of Ctrl and treated cells was

653 indicated by asterisks \*\*\* (P < 0.001); \*\* (P < 0.01) and \* (P < 0.05) (One-way

654 ANOVA with Dunnett's multiple comparison test).

655

656 **Figure 3.** Progesterone release (%) by NCI-H295R cell line in culture after 48 h of

657 NiCl<sub>2</sub> exposure. *Abbreviations:* Each point represents the arithmetic mean ( $\pm$ S.D.)

658 progesterone % of (untreated) controls (Ctrl) determined of three repeated experiments.

659 The number of replicate wells was 4-6 at each point per experiment. The statistical

660 difference between the values of Ctrl and treated cells was indicated by asterisks \*\*\* (P

661 < 0.001); \*\* (P < 0.01) and \* (P < 0.05) (One-way ANOVA with Dunnett's multiple

662 comparison test).

663

664 **Figure 4.** Testosterone release (%) by NCI-H295R cell line in culture after 48 h of

665 NiCl<sub>2</sub> exposure. *Abbreviations:* Each point represents the arithmetic mean ( $\pm$ S.D.)

666 testosterone % of (untreated) controls (Ctrl) determined of three repeated experiments.

667 The number of replicate wells was 6-10 at each point per experiment. The statistical

668 difference between the values of Ctrl and treated cells was indicated by asterisks \*\*\* (P

669 < 0.001); \*\* (P < 0.01) and \* (P < 0.05) (One-way ANOVA with Dunnett's multiple  
670 comparison test).

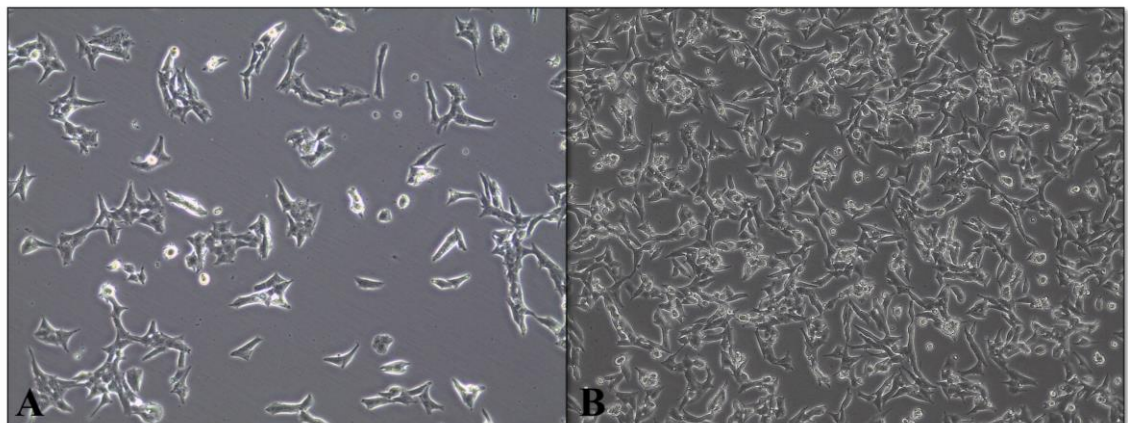
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672 **Figure 5.** 17 $\beta$ -estradiol release (%) by NCI-H295R cell line in culture after 48 h of  
673 nickel chloride (NiCl<sub>2</sub>) exposure. *Abbreviations:* Each point represents the arithmetic  
674 mean ( $\pm$ S.D.) 17 $\beta$ -estradiol % of (untreated) controls (Ctrl) determined of three repeated  
675 experiments. The number of replicate wells was 6-12 at each point per experiment. No  
676 statistical difference between the values of Ctrl and treated cells was not recorded (P >  
677 0.05) (One-way ANOVA with Dunnett's multiple comparison test).

678

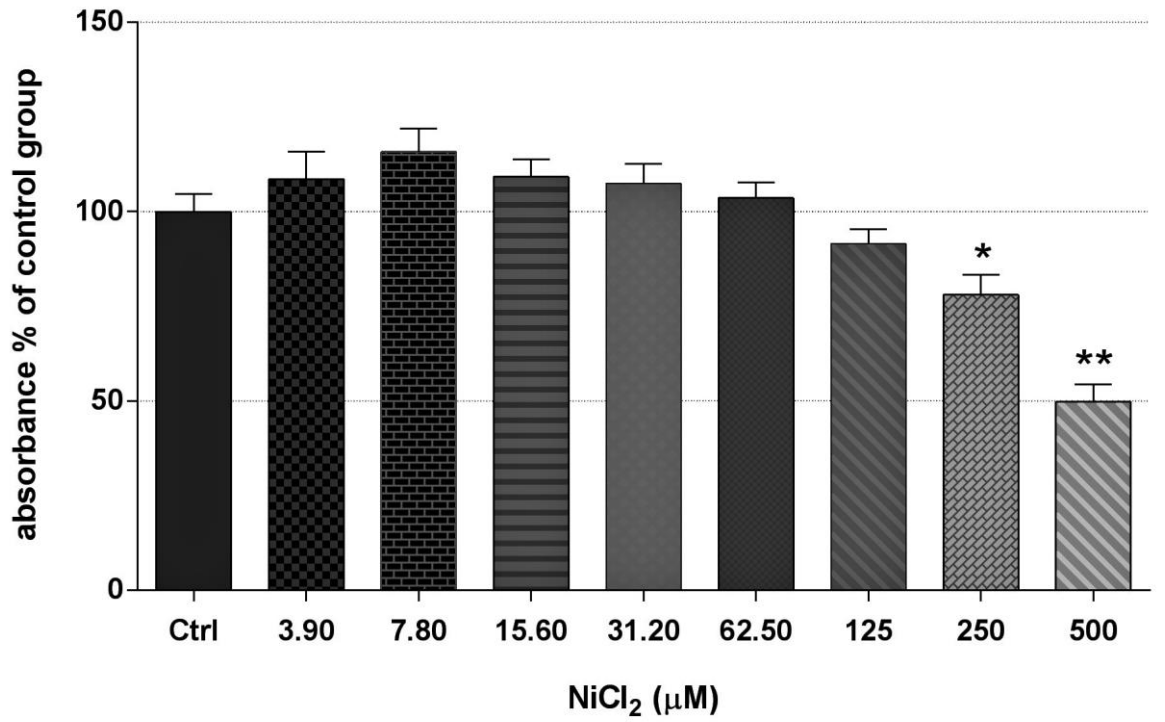
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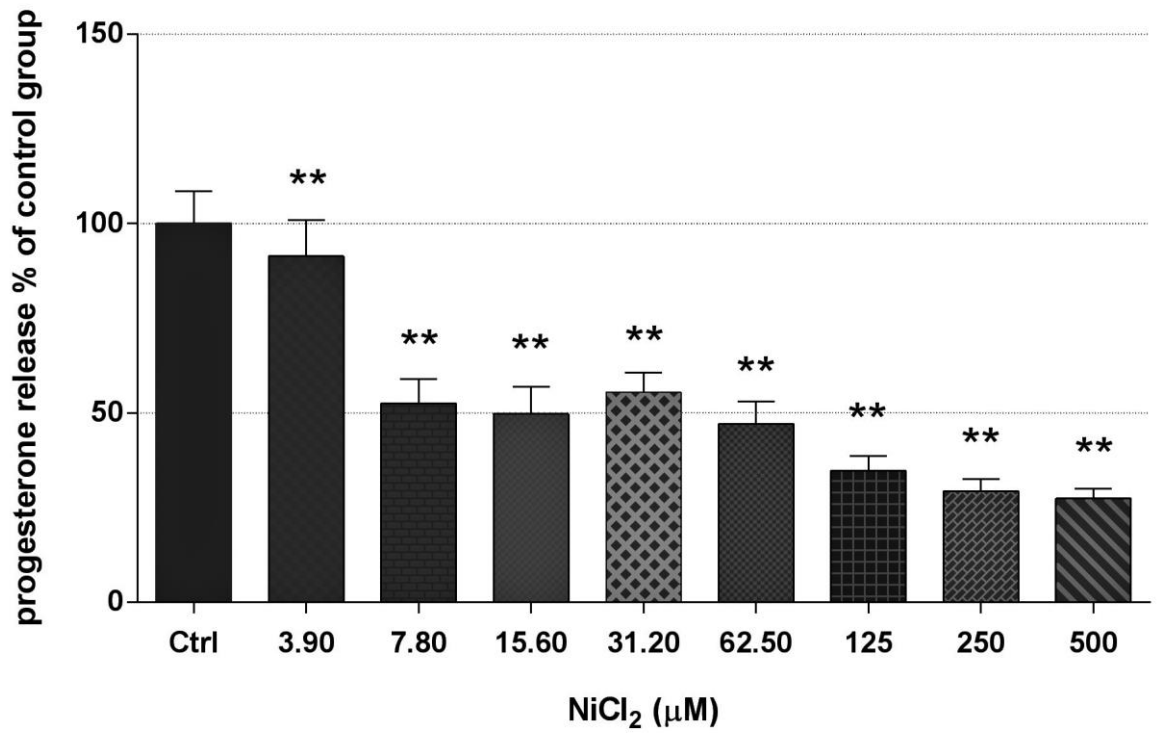
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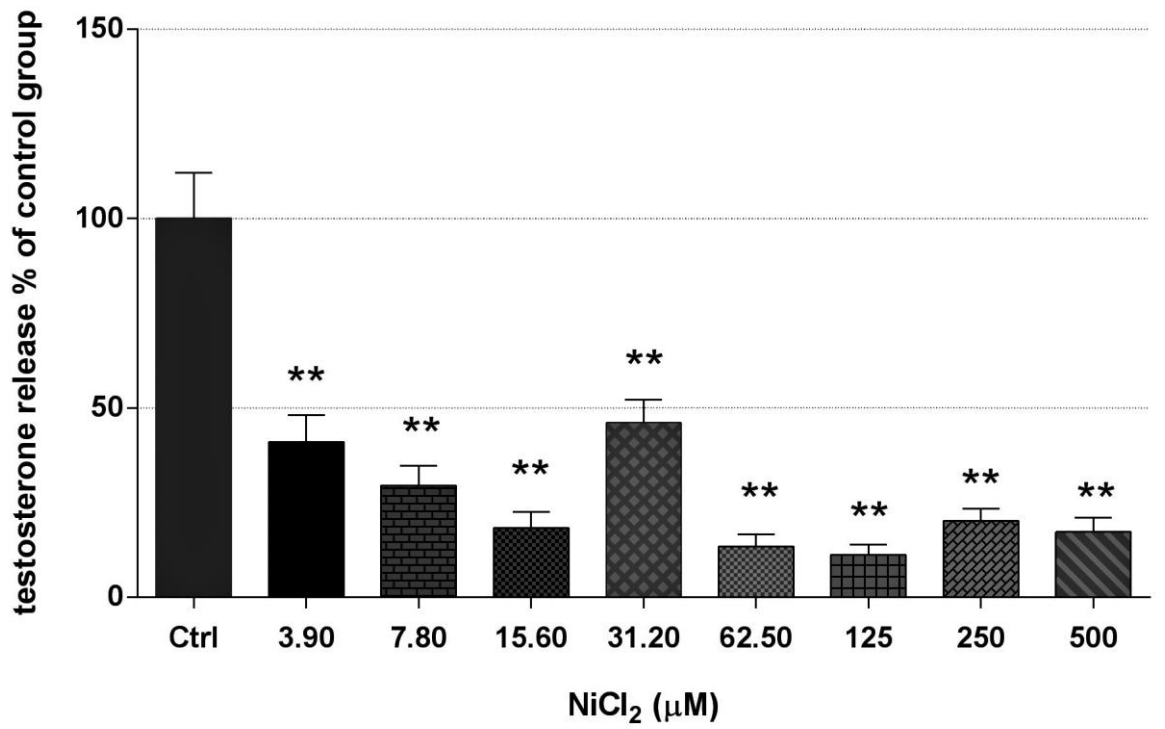


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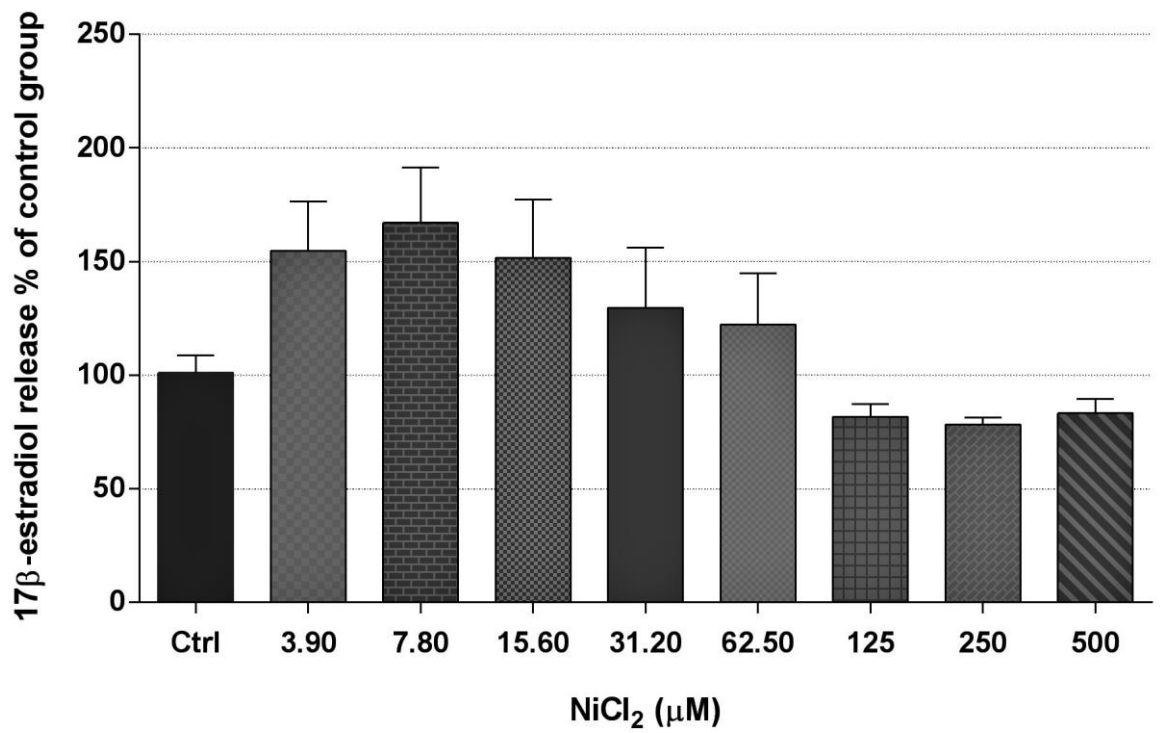
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